

Corchorusolitoriuswaste(mulukiya) as a potential sorbent for the removal of cadmium (II) andthorium (IV) ions: kinetics, equilibrium study and image analysis

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Abstract: This work was conducted to determine the practicability of using a new adsorbent *Corchorusolitorius(mulukiya)*waste,for the removal of cadmium (Cd(II)), and thorium (Th(IV)) from wastewater. *Corchorusolitorius*wereanalysis by Fourier transform infraredFTIR , scanning electron microscopy (SEM) and energy dispersive X-ray Spectroscopy (EDX). Some parameters such as adsorbent dosage, solution pH's, initial metal ion concentrations, and contact time, that influence adsorption phenomenon, were studied. The optimum pH for maximum adsorption of Cd(II) and Th(IV) was found to be 5.55 and 4.50, respectively. The contact time required for reaching equilibrium was 2 hr. The pseudo second-order kinetic model was the best fit to represent the kinetic data. Analysis of the equilibrium adsorption data using Langmuir and Freundlich models showed that theLangmuir model was well suitable to describe the metal ions adsorption.

Keywords:*Corchorusolitorius*waste; Cadmium(II); Thorium(IV); Adsorption.

I. Introduction

The discharge of radioactive waste into the environment is a major issue. A radioactive material, is an environmental pollutant, being present in mining industry wastewater. Other sources include mainly soils and seawater (Rao et al., 2006).Thorium is abundantly available radioactive element in earth's crust in association with rare earths and uranium. The main sources of thorium are monazite, rutile and thorianite. It occurs in the tetravalent form in its compounds in nature. Also occurs naturally in ocean. The recent attention to extract thorium from waste streams is due to its application as a good fertile nuclear fuel for breeder reactors. It is extensively used in variable areas, such as optics, radio, aeronautics and aerospace, metallurgy and chemical industry, nuclear industry and material fields (Zhang et al., 2005).The existence of thorium in the nature and its potential use in the nuclear technology were not considered with sufficient importance because of the geological availability of natural resources of thorium and uranium. Thorium is an important component of the fuel in nuclear breeder reactors and thorium fuel cycle can be used in most of the reactors already operated (Raje and Reddy, 2010; Unak, 2000). On the verge of energy crisis, world has been utilizing maximum of its nuclear energy resources by extracting radioactive metals like uranium, thorium etc. Nuclear waste at a thorium-based reactor is minimal, and the plutonium from the process is used for further generation (Unak, 2000). Since thorium ions are radioactive, they emit electromagnetic radiations that can give permanent damage to cells and ultimately resulting to cancer.

Heavy metals are often detected in industrial wastewaters, which originate from metal plating, mining activities, smelting, battery manufacture, tanneries, petroleum refining, paint manufacture, pesticides, pigment manufacture, printing and photographic industries, etc. (Kadirvelu et al,2001, Williams et al,1998). Presence of metal ions is of specialconcern as they can accumulate in different components of the environment; represent a serious threat to human populations, the fauna, etc. Cd(II) is one of the most toxic metals. The cadmium has been well recognized for its negative effect on the environment where it is non-degradable and accumulates readily in living systems.Adverse health effects due to cadmium are well documented and it has been reported to cause renal disturbances, lung insufficiency, bone lesions, cancer and hypertension in humans (Parker, 1980).

A number of technologies for removing cadmium and thorium ions from wastewaterhave been developed. The most important of these techniques include chemicalprecipitation, solvent extraction, adsorption, ion-exchange, and reverse osmosis. Adsorption is a relatively simple, cheap and effective method for wastewater treatment.Different adsorbents have been reported in the literature for the removal of U(VI) and Th(IV) ions, such as cellulose triacetate (Villalobos-Rodríguez et al., 2012), magnetic chitosan composite (Hritcu et al., 2012) and sepiolites (Kilisliloglu and Aras, 2010).

A special attention has been focused on the use of natural adsorbents as an alternative to replace the conventional adsorbents, based on both the environmental and the economical points of view (Bable, 2003). Natural materials that are available in large quantities, or certain waste products from industrial or agricultural

operations, may have potential as inexpensive sorbents. Due to their low cost, when these materials the end of their lifetime, they can be disposed of without expensive regeneration. The abundance and availability of agricultural by-products make them good sources of raw materials for natural sorbents.

Corchorusolitorius Linn is an important fiber crop cultivated in the Gangetic plane of West Bengal, India and Egypt. Jute plants produce edible leaves that are used as a vegetable and food constituent common to the people. It is made into a slimy soup or sauce in some of West African cooking traditions. Furthermore, it is also a popular dish in the northern provinces of the Philippines, also known as saluyot (Oboh et al., 2012).

The objective of the present work was to study the adsorption characteristics of cadmium (Cd(II)), and thorium (Th(IV)) ions from aqueous solution using *Corchorusolitorius* (mulukiya). The Langmuir and Freundlich and models were used to describe equilibrium isotherms. The kinetic data were fitted by models including pseudo-first-order and pseudo-second-order.

II. Materials and methods

2.1 Materials

All chemicals and reagents needed in the present study were of analytical grade and purchased from Merck. Solutions of Cadmium and thorium were prepared from $\text{Cd}(\text{NO}_3)_2 \cdot 5\text{H}_2\text{O}$ and $\text{Th}(\text{NO}_3)_4 \cdot 5\text{H}_2\text{O}$ by dissolving appropriate amounts of salts in double distilled water. The stock solution was diluted with distilled water to obtain the desired concentration.

Corchorusolitorius (Fig.1) were collected from Elsharkia city (Egypt). The raw materials (stems of *Corchorusolitorius*) were washed thoroughly with deionized water, and dried at 100 °C, crushed and sieved to particles with size range of 40–60 mesh for use as adsorbent



Fig. 1– Photo of *Corchorusolitorius* leaves.

The *Corchorusolitorius* belongs to the class of condensed tannins which are the polymerized products of flavan-3-ols and/or flavan-3,4-diols and are the polyphenols with molecular weight of 500–3000 Da (Pizzi, 1993). The branch-chained tetraflavonoid structures of *Corchorusolitorius* have reactive and accessible nucleophilic centers in the rings (Fig. 2) (Aparna, et al (2013)).

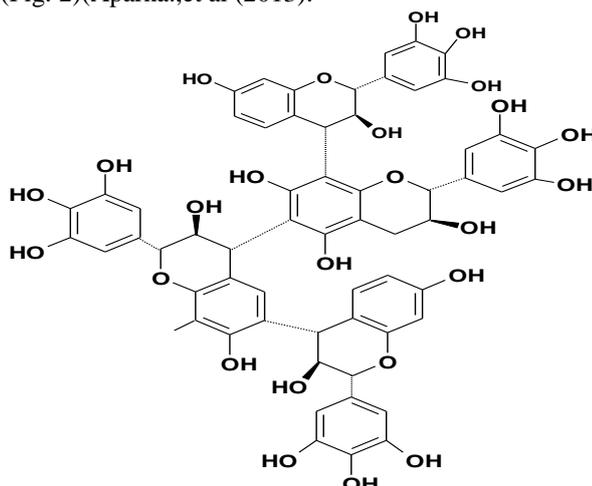


Fig. 2. Branch-chained tetraflavonoid structure of *Corchorusolitorius* unit.

2.2 Analysis

The FT-IR spectra of the sample were taken using a Shimadzu FT-IR spectrophotometer between 4000 and 500 cm^{-1} with KBr background. The *Corchorusolitorius* was done using Thermo Finnigan Flash EA III Z CHNS analyzer.

The surface morphology of Corchorusolitorius and Cd(II) and Th(IV) adsorbed onto Corchorusolitorius samples examined using (JEOL SEM – 25) scanning electron microscope. Prior to examination, the samples were dried under sputter coated gold.

The energy dispersive X-ray (EDEX) pattern of Corchorusolitorius and Cd(II) and Th(IV) adsorbed onto Corchorusolitorius using Oxford energy dispersive X-ray spectroscopy.

2.3 Adsorption experiments

Glass-stoppered Erlenmeyer flasks were used for adsorption experiments. Batch mode adsorption studies were conducted to study the effects of dose, equilibrium time, initial metal ion concentrations, and pH. All experiments were carried out at 25°C. A known amount of adsorbent was continuously shaken in 50 mL of the corresponding salt solutions. Shaking was kept at 300 r/min for 2hr to reach equilibrium. After attaining the equilibrium, the adsorbent was removed by filtration and the concentrations of Cd(II) and Th(IV) ions in the corresponding filtrate were determined.

2.4 Metal analysis

The concentration of Cd(II) and Th(IV) ions was determined by AAS method. The difference in concentrations was taken as the amount adsorbed by Corchorusolitorius. An atomic adsorption spectrometer (Solar M6, USA), with lead hollow cathode lamp and air acetylene flame.

The amount of metal ions adsorbed at equilibrium, q_e (mg/g), was obtained by Eq. (1):

$$q_e = \frac{(C_0 - C_e) \times V}{m} \quad (1)$$

where, C_0 (mg/L), C_e (mg/L) are the initial and equilibrium concentrations of the metal ion respectively, V (L) is the volume of solution, and m (g) is the mass of the adsorbent. Each experiment was carried out in duplicate and the average value is reported.

Solution pH was measured with a pH Meter (model μ -362). All the solutions, for the adsorption equilibrium studies, were shaken at 25°C controlled by water bath shaker purchased from Labline, India.

III. Results and discussion

3.1 Characterization of adsorbent:

3.1.1 Energy dispersive X-ray (EDEX) analysis

The elucidation of Cd(II) and Th(IV) adsorption on the Corchorusolitorius were done by energy dispersive X-ray (EDEX) analysis. (EDEX) spectrum of Corchorusolitorius before and after interaction with metal ions is shown Fig.3(a) (b) and (c). The spectrum of adsorbent presented in Figure 3(a) and Table (1) show that the major elements present. The EDX spectrum for Corchorusolitorius particles indicated the presence of some element in the Table (1) but did not show the characteristic signal of Cd(II) and Th(IV) ions on the surface. The interaction between Corchorusolitorius with Cd(II) and Th(IV) are depicted in Fig.3 (b) and (c). An additional signal compared with the control ones Fig.3 (a) observed in EDEX profile (b) and (c) suggesting the interaction of metal with functional groups present in adsorbent. The EDX spectrum gives the characteristic peaks for Cd(II) and Th(IV). This confirms the binding of the metal ions to the Corchorusolitorius surface. The spectra reflected Th(IV) with a high adsorption than Cd(II).

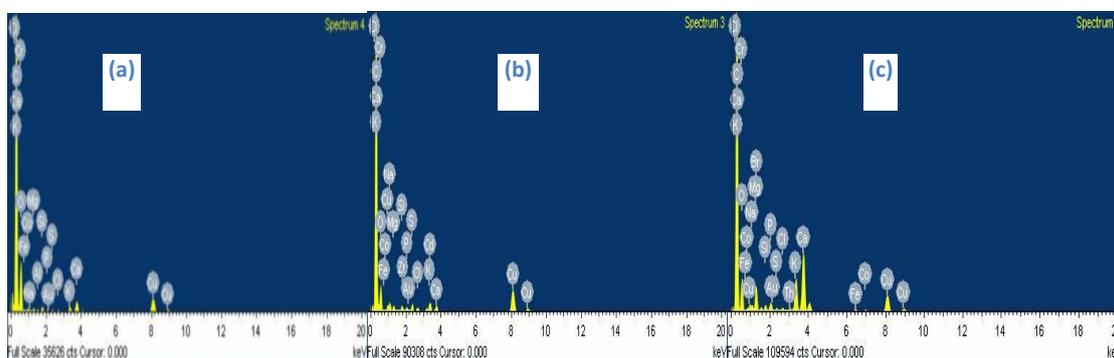


Fig. 3. EDEX Spectra of (a) Corchorusolitorius (b) Cd (II)-Corchorusolitorius (c) Th (IV) Corchorusolitorius.

Table1. Weight % of (a) Corchorusolitorius (b) Corchorusolitorius with Cd(II) (c) Corchorusolitorius with Th(IV).

Corchorusolitorius		Cd-Corchorusolitorius		Th-Corchorusolitorius	
Element	Weight%	Element	Weight%	Element	Weight%
C	22.73	C	20.64	C	19.07
N	1.45	Na	0.22	Na	0.20
Na	0.08	Mg	0.39	Mg	1.26
Mg	0.19	Al	0.17	Si	0.28
Al	0.05	Si	3.36	P	0.62
Si	0.10	P	0.15	S	0.23
P	0.23	S	0.28	Cl	0.20
S	0.37	Cl	0.17	K	3.75
Cl	0.18	K	1.06	Ca	7.22
K	0.39	Ca	1.18	Cr	0.09
Ca	0.80	Cr	0.06	Fe	0.37
Cr	0.05	Fe	0.40	Co	0.24
Fe	0.08	Cu	8.28	Cu	6.75
Cu	5.53	Cd	0.59	Br	0.24
Au	0.07	Au	0.10	Au	0.04
O	67.69	O	62.94	Th	0.60
Totals	100.00	Totals	100.00	O	58.83
				Totals	100.00

3.1.2 Scanning electron microscopic(SEM) analysis

The scanning electron micrographs (SEM) of Corchorusolitorius before and after adsorption Cd(II) and Th(IV) are shown in (Fig. 4(a),(b) and (c)). The surface morphology revealed that a significant change in the surface topography of Corchorusolitorius after metal adsorption. The surface of raw Corchorusolitorius was smooth with multicellular nature, whereas rough surface morphology with fragments and groove like structures was observed due to adsorbed of Cd(II) and Th(IV). Moreover, it was also observed that the natural green color of Corchorusolitorius had become dark green after adsorption.

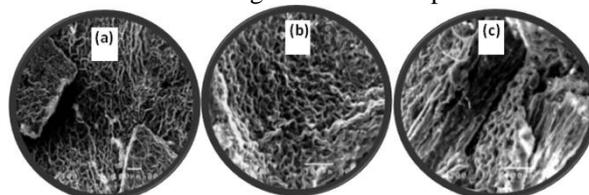


Fig. 4. SEM images of (a) Corchorusolitorius (b) Cd(II)-Corchorusolitorius (c) Th(IV) Corchorusolitorius.

3.1.3 Fourier transform infrared (FTIR) spectroscopy

Fig.5(a),(b) and (c) presents the FT-IR spectra of Corchorusolitorius, Cd(II)-Corchorusolitorius and Th(IV)- Corchorusolitorius respectively. The FTIR spectra of a,b,c, showed a broad band in the region of 3700–3010 cm^{-1} , which is due to the characteristic –OH stretching for phenolic or alcoholic group. The smaller peak at 2927 cm^{-1} is associated with the –CH stretching of the aromatic rings and methylene(–CH₂–) bridges. The peak at 1611 cm^{-1} corresponds to the characteristic elongation of aromatic –C=C– bonds. The deformation vibration of –C–C– bonds in the phenolic group absorbs in the region of 1510–1400 cm^{-1} . The peak at 1454 cm^{-1} may be assigned to bending. The signal at 1345 cm^{-1} is associated with the –OH deformation vibration for phenolic or alcoholic group. The absorption peak at around 1232 cm^{-1} is associated with the –CO stretching of aromatic ring. The ant symmetric –C–O–C– stretching contributes for the peak at 1159 cm^{-1} . The bands in the range of 843–665 cm^{-1} are attributed to the deformation vibrations of the C–H bond in the aromatic rings (Sanchez-MartGnaGonzalez-Velasco, Beltran-Heredia, Gragera-Carvajal, & Salguero-Fernandez, 2010). The peak at 898 cm^{-1} for β -glucosidic linkage of cellulose in Corchorusolitorius. It was also observed that new peak at 556 cm^{-1} due to adsorbed of Cd(II) and Th(IV) onto Corchorusolitorius (Fig. 5(b) and (c)).

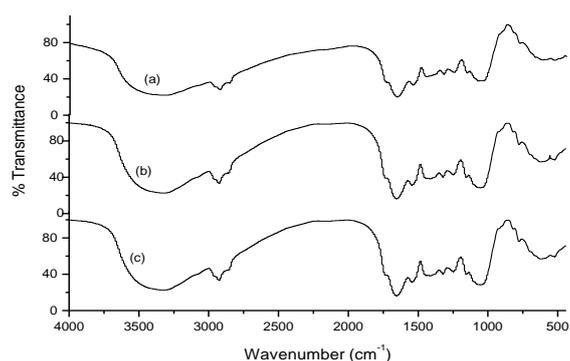


Fig. 5. FTIR spectra of (a) *Corchorusolitorius* (b) Cd(II)-*Corchorusolitorius* (c) Th(IV) *Corchorusolitorius*.

3.2 Effect of pH

The pH level is one of the important parameters for the adsorption of metal ions from an aqueous solution. The pH of the solution will affect the surface charge of the adsorbent and the degree of ionization. The effect of pH on the adsorption of Cd(II) and Th(IV) onto *Corchorusolitorius* at different initial concentrations (30.0, 50.0, 70.0 and 100.0 mg/L) at 25°C is depicted in Fig. 6. It can be seen that the adsorption percentages are very low at strong acidic medium. The minimal adsorption at low pH may be due to the higher concentration and high mobility of H⁺, which are preferentially adsorbed rather than metal ions (Ajmal et al., 2000, Annadurai et al., 2002). Increasing the pH of the solution, the lower number of H⁺ (results in more metal ions adsorption). Adsorption of increased sharply with increase of pH up to optimum pH 5.55 and 4.50 for Cd(II) and Th(IV) respectively then started decreasing. Therefore, all the following experiments were performed at these optimum pH. The higher adsorption Th(IV) at pH 4.50 could be associated with the predominance of [Th₂(OH)₂]⁶⁺ and other polymerized species possessing a better binding affinity (Tsezos and Volesky, 1982). Adsorption was limited above pH 6.0 due to the formation of neutral Th(OH)₄.

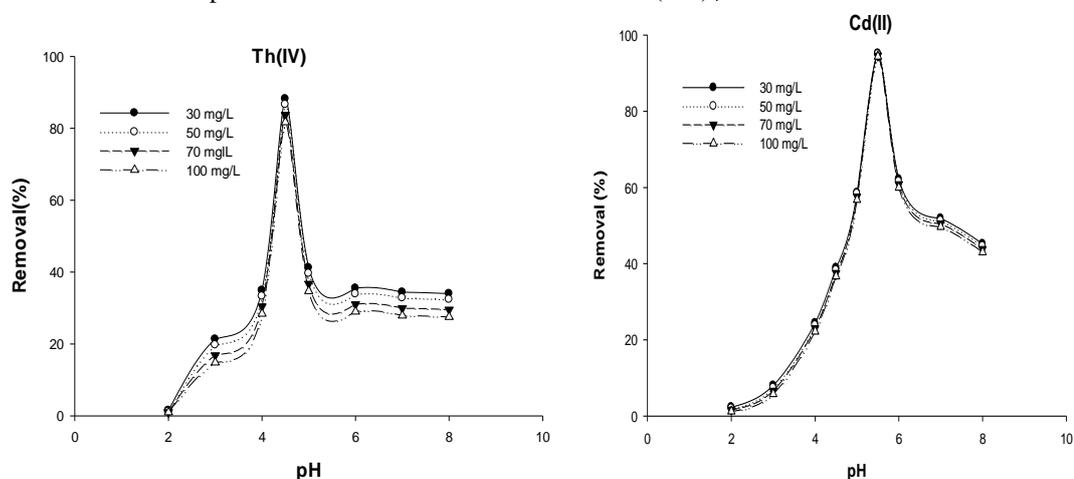


Fig. 6 Effect of pH on the adsorption of Cd(II) and Th(IV) onto *Corchorusolitorius*. Conditions: adsorbent dose 0.01 g/L; equilibrium time 4 hr; temperature 25°C.

3.3 Effect of adsorbent dose

The effect of adsorbent dose on the removal of Cd(II) and Th(IV) ions was studied by carrying out the adsorption experiments using various amounts of *Corchorusolitorius* at the optimum pH 5.55 for Cd(II) and 4.5 for Th(IV) at 25°C with varying initial metal ion concentration of Cd(II) and Th(IV) 30, 50, 70 and 100 mg/L Th(IV) (Fig. 7). The time needed to attain equilibrium was taken as 4 hr. The removal percentage of Cd(II) and Th(IV) ions was noted as a function of adsorbent dose. The removal of metal ion increased with an increase of dose from 0.005 to 0.03 g/L. This can be due to the fact that, the active sites could be effectively utilized when the dosage was low. As adsorbent dosage was increased, the number of unoccupied effective sites and the surface area also increased and thus it improved the percentage of adsorption.

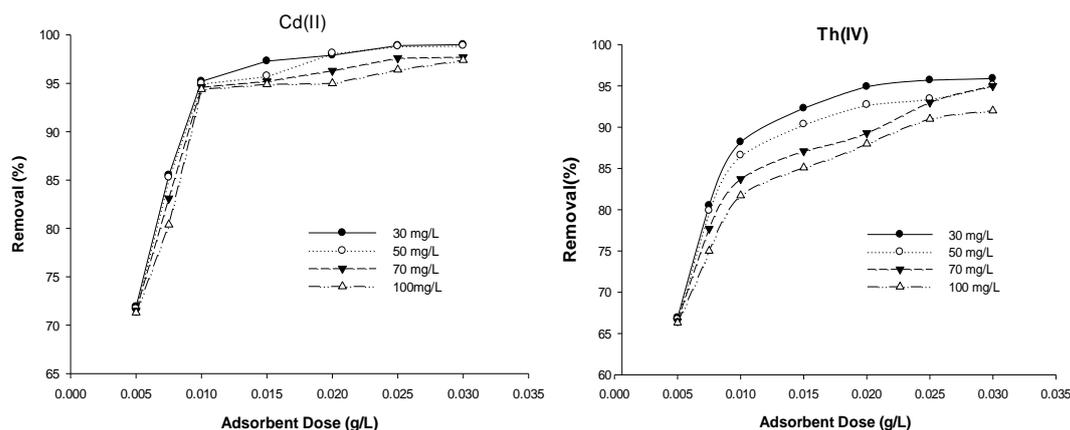


Fig.7Effect of adsorbent dose on the adsorption of Cd(II) and Th(IV) onto *Corchorusolitorius*. Conditions: equilibrium time 2 hr; concentrations; 30, 50,70, and 100 mg/L; pH 5.0 for Cd(II); pH 4.5 for Th(IV).

3.4 Effect of contact time and initial concentration

In order to establish an appropriate contact time between the adsorbent and metallic ions in solution, the adsorption of Cd(II) and Th(IV) onto *Corchorusolitorius* was tested as a function of contact time at different initial concentrations (30 to 100mg/L) of metal ion solutions . Fig. 8 illustrated that the rate of adsorption for Cd(II) and Th(IV) was initially fast and then reached a plateau value. The high adsorption rate was due to the abundance of active binding sites on the adsorbent, and as these sites get saturated, only a reduced amount of adsorption was occurred (da Costa and Leite, 1991). The equilibrium time was found to be a constant and was independent of the initial Cd(II) and Th(IV) ion concentrations.

The quantity of Cd(II) and Th(IV) ions adsorbed at equilibrium onto *Corchorusolitorius* was found to be improved from 26.5 and 31.09 mg/g to 70.42 and 73.2 mg/g, respectively, on increasing the initial concentration from 30 to 100 mg/L indicating that the metal removal was concentration dependent. The increase of metal ion concentration accelerated the diffusion of metal ion from the bulk solution onto the adsorbent phase because of the enhanced force of the higher concentration (Hameed et al., 2008). If metal ion/adsorbent ratios are low, metal adsorption takes place at higher energy sites. When the metal ion/adsorbent ratio is increased, the higher energy sites become saturated and lower energy sites are involved in adsorption and consequently the adsorption efficiency is decreased (Bhattacharya et al., 2006). Since the amount of metal ions adsorbed did not change after 2 hr, the succeeding adsorption experiments were done keeping 2hr as an appropriate equilibrium time.

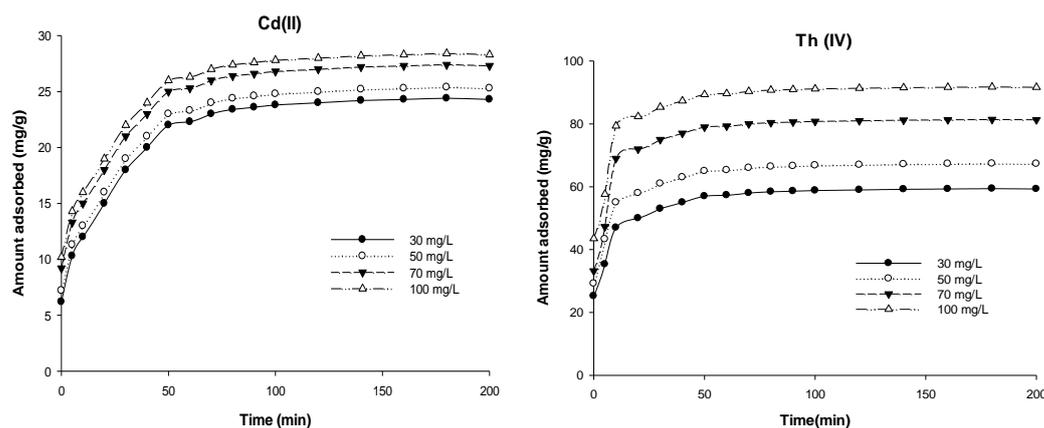


Fig. 8 Adsorption kinetics at different initial Cd(II) and Th(IV) concentrations (30–100 mg/L). Conditions: adsorbent dose 0.01 g/L; pH 4.5 for Th(IV); pH 5.0 for Cd(II); temperature 25⁰C.

3.5 Adsorption kinetics

Studies on the adsorption rate showed that the removal decreased with increased Cd(II) and Th(IV) concentration Fig. 8. The metal removal mechanism is related to the initial metal ion concentration. Adsorption of metal ions takes place at definite sites only when the metal ion concentration is low, but as the concentration is increased, the sites become saturated and exchange sites are filled (Patel et al., 2007). The maximum amount of Cd(II) and Th(IV) ions was adsorbed within the first 30 min (80–90% of total metal ions adsorbed) and

thereafter the adsorption proceeded at a slower rate until equilibrium reached. The equilibrium time was found to be at 120 min for the initial concentration range studied. This might be attributed to extremely slow diffusion of the metal ions from the surface film into the microspores which are the least accessible sites for adsorption (Warhurst et al., 1997).

The kinetics of adsorption describes the rate of metal ions uptake onto the *Corchorusolitorius* and this rate control the equilibrium time. The kinetics of adsorbate uptake is required for selecting optimum operating conditions for the full-scale batch process, so these models are important in water treatment process design. The kinetic parameter, which is helpful for the prediction of adsorption rate, gives important information for designing and modeling the process. Thus, the effects of contact time at different concentration were analyzed from the kinetic point of view.

Adsorption kinetics are generally controlled by different mechanism, of which the most limiting are the diffusion mechanisms, including the initial curved portion, attributed to rapid external diffusion or boundary layer diffusion and surface adsorption, and the linear portion, a gradual adsorption stage due to the intraparticle diffusion, followed by a plateau to the equilibrium where the intraparticle diffusion starts to decrease due to the low concentration in solution phase as well as fewer available adsorption sites (Guibalet al., 2003). Previously several researchers used different kinetic models, such as Lagergren's pseudo first order, pseudo second order, Elovich kinetic equation, and parabolic diffusion model, to represent the mechanism of the adsorption process (Hoda et al., 2006; Sarkar et al., 2006; Weber and Morris, 1963).

The Lagergren's pseudo first order and pseudo-second-order model has been widely used for adsorption systems due to its good representation of the experimental data for most of the adsorbent-adsorbate systems (Ho and McKay, 1999). The pseudo-second-order equation has the following advantages: it does not have the problem of assigning an effective adsorption capacity, the adsorption capacity, rate constant of pseudo-second order and the initial adsorption rate all can be determined from the equation without knowing any parameter beforehand.

The linear form of pseudo first order model can be expressed as:

$$\log(q_e - q_t) = \log q_e - \frac{k_1}{2.303} t \quad (2)$$

where q_e and q_t (mg g^{-1}) are the adsorption capacities at equilibrium and at time t respectively

k_1 at different concentration evaluated experimentally, were calculated using the slope and intercept of plots of $\log(q_e - q_t)$ versus t , Table 2. Best fit lines at each concentration yielded relatively high R^2 values.

The pseudo second-order rate expression, which has been applied for analyzing sorption kinetics from liquid solutions, is linearly expressed as

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{1}{q_e} t \quad (3)$$

where k_2 is the rate constant for pseudo second-order adsorption ($\text{mg}^{-1} \text{min}^{-1}$) and $k_2 q_e^2$ or h is the initial sorption rate ($\text{mg}^{-1} \text{min}^{-1}$). k_2 and q_e can be obtained from the intercept and slope of plotting (t/q_t) versus (t) . where h is the initial sorption rate ($\text{mg}^{-1} \text{min}^{-1}$); k_2 ($\text{mg}^{-1} \text{min}^{-1}$) is the second-order rate constant. k_2 and q_e can be obtained from the intercept and slope of plotting (t/q_t) versus (t) . This model is more likely to predict the kinetic behavior of adsorption with chemical adsorption being the rate controlling step. The pseudo-second-order reaction is greatly influenced by the amount of metal on the adsorbent's surface and the amount of metal adsorbed at equilibrium (Ho and McKay, 1999).

Pseudo second-order adsorption parameter k_2 were determined at different concentration and recorded in Table 2. Furthermore the correlation coefficients R^2 for the pseudo second-order kinetic model fits are close 1.00, higher than the correlation coefficients derived from pseudo first-order model fits. Given the good agreement between model fit and experimentally observed equilibrium adsorption capacity in addition to the large correlation coefficients, this suggests that Cd(II) and Th(IV) adsorption followed pseudo second-order kinetics and adsorbent ions were adsorbed onto the *Corchorusolitorius* via chemical interaction.

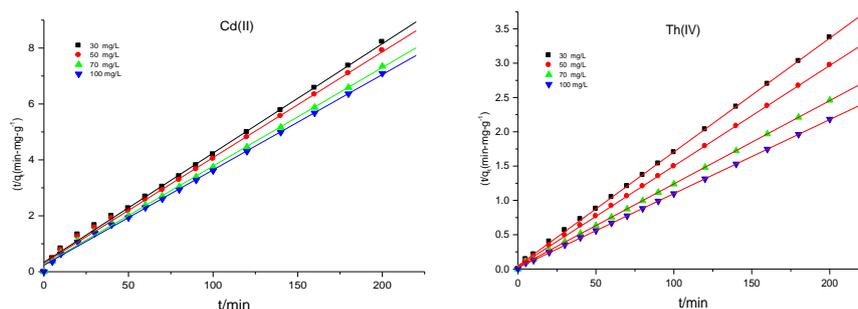


Fig. 9 Pseudo-second-order kinetics model for Cd(II) and Th(IV) concentrations (30–100 mg/L) adsorption onto the *Corchorusolitorius*.

Table 2 Kinetic parameters for the adsorption of Cd(II) and Th(IV) onto *Corchorusolitorius*

Concentration (mg/L)	Pseudo first -order		Pseudo second -order		
	k_1 ($\times 10^{-3} \text{ min}^{-1}$)	R^2	k_2 ($\times 10^{-3} \text{ g}/(\text{mg}\cdot\text{min})$)	R^2	
Cd(II)	30	8.86	0.993	4.59	0.998
	50	7.82	0.994	4.72	0.999
	70	7.59	0.992	4.64	0.999
	100	8.28	0.995	5.06	0.999
Th(IV)	30	8.88	0.984	5.73	0.999
	50	8.94	0.982	5.92	0.999
	70	9.24	0.970	5.47	0.999
	100	9.93	0.974	5.64	0.999

3.5 Adsorption isotherm

The adsorption data are usually collected in terms of the total amount of the adsorbate adsorbed at the adsorbent’s surface. When the amount adsorbed at equilibrium is plotted against the equilibrium concentration of the adsorbate at constant temperature, adsorption isotherms are obtained. Since an adsorption isotherm gives an idea about the adsorption capacity of the adsorbent, this can be used for optimizing the application of an adsorbent. The common adsorption isotherm models are Langmuir and Freundlich, (Langmuir-Freundlich). These equations were tested to fit the experimental adsorption data of Cd(II) and Th(IV) ions. The Langmuir model suggests a monolayer adsorption onto the adsorbent having uniform surface sites with equivalent energies and the adsorbed species do not interact with one another (Langmuir, 1918). The Freundlich model is an empirical one relating adsorption onto heterogeneous surface (Freundlich, 1906). This model is typical of heterogeneous surfaces and suggests that the surface active sites of the adsorbent have unequal binding energies.

Langmuir equation:

$$q_e = \frac{Q^0 b_L C_0}{1 + b_L C_0} \quad (4)$$

Freundlich equation:

$$q_e = K_F C_e^{1/n_F} \quad (5)$$

where, q_e (mg/g) is the amount adsorbed at equilibrium and C_0 (mg/L) is the equilibrium concentration of the metal ions in solution. Q^0 and b_L are Langmuir constants associated with adsorption capacity and binding energy of adsorption, respectively. K_F and $1/n_F$ represent Freundlich’s constants connected to adsorption ability and the heterogeneity factor, respectively. The isotherms constants of Cd(II) and Th(IV) ions onto *Corchorusolitorius* are illustrated in (Table 3). The q_e , was studied as a function of C_0 . Based on the correlation coefficients (Table 3), the Langmuir equation gives a better fit of experimental data than the Freundlich, for the adsorption of Cd(II) and Th(IV) ions. This suggests a monolayer adsorption onto the adsorbent having uniform surface sites with equivalent energies, The Freundlich constant $1/n_F$ was below 1 at all the experimental concentrations, indicative of high adsorption intensity (Tsai et al., 2003).

Table 3 Various isotherm parameters for the adsorption of Cd(II) and Th(IV) onto Corchorusolitorius.

Metal ion	Langmuir	Freundlich		
Cd(II)	Q^0	95.55	K_F	4.58
	b_L	0.0121	$1/n_F$	0.007
	R^2	0.960	R^2	0.9568
Th(IV)	Q^0	44.94	K_F	2.79
	b_L	0.043	$1/n_F$	0.0372
	R^2	0.979	R^2	0.9755

IV. Conclusions

Batch adsorption studies proved that the new adsorbent, Corchorusolitorius has high adsorption capacity for Cd(II) and Th(IV) ions in the pH range 4.0–6.0. Kinetics data fitted well with pseudo second-order model confirming the chemisorptions of the metal ions onto Corchorusolitorius. The adsorption equilibrium experiments matched well with the Freundlich isotherm model. The results of this work showed that the new adsorbent Corchorusolitorius can be effectively used for the removal of metal ions from aqueous solutions and other industrial effluents.

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